Florida Keys National Marine Sanctuary

Water Quality Monitoring Project

1999 Annual Report

Principal Investigators

Ronald D. Jones

and

Joseph N. Boyer

Southeast Environmental Research Center
Florida International University
Miami, FL 33199

US EPA Contract #X994621-94-0

Table of Contents

I. Executive Summary

This report serves as a summary of our efforts to date in the execution of the water quality monitoring project for the FKNMS. Since initiation we have added 4 sampling sites and adjusted 6 others to increase cover in the Sanctuary Preservation Areas and Ecological Reserves. We have received 21 requests for data by outside researchers working in the FKNMS of which one has resulted in a master's thesis. Two scientific manuscripts have been submitted for publication: one is a book chapter in Linkages Between Ecosystems: the South Florida Hydroscape, St. Lucie Press; the other is in special issue of Estuarine, Coastal and Shelf Science on visualization in coastal marine science. Two other manuscripts are being prepared; one in conjunction with the FKNMS seagrass monitoring program. We maintain a website where data from the FKNMS is integrated with the other parts of the SERC water quality network (Florida Bay, Whitewater Bay, Biscayne Bay, Ten Thousand Islands, and SW Florida Shelf) and displayed as downloadable contour maps - http://www.fiu.edu/~serc/jrpp/wqmn/datamaps/datamaps.html

The period of record for this report is Mar. 1995 - Oct. 1999 and includes data from 17 quarterly sampling events at 154 stations within the FKNMS including the Dry Tortugas National Park. Field parameters at each station include salinity, temperature, dissolved oxygen (DO), turbidity, relative fluorescence, and light attenuation (K_d). Water chemistry variables measured at each station include the dissolved nutrients nitrate (NO_3^-), nitrite (NO_2^-), ammonium (NH_4^+), dissolved inorganic nitrogen (DIN), and soluble reactive phosphate (SRP as PO_4^{3-}). Total unfiltered concentrations of organic nitrogen (TON), organic carbon (TOC), phosphorus (TP), and silicate ($Si(OH)_4$) were also measured. The monitored biological parameters included chlorophyll *a* (Chl *a*) and alkaline phosphatase activity (APA).

Grouping stations by depth showed that temperature, DO, TOC, and TON were generally higher at the surface while salinity, NO_3^-, NO_2^-, NH_4^+, TP, and turbidity were higher in bottom waters. This slight stratification is indicative of a weak pycnocline which is maintained by freshwater inputs and solar heating at the surface. Elevated nutrients in the bottom waters is due to benthic flux and some upwelling. Stations grouped by to geographical region showed that the Tortugas and the Upper Keys had lower nutrient concentrations than the Middle Keys or Lower Keys. In the Lower Keys DIN was elevated in the Backcountry. TP concentrations in the Lower

Keys transects decreased with distance offshore but increased along transects in the Upper Keys, mostly because of low concentrations alongshore. The Sluiceway had lowest salinity and highest TOC, TON, and $Si(OH)_4$ concentrations. The north Marquesas area exhibited highest phytoplankton biomass for any segment of the FKNMS. Declining inshore to offshore trends were observed for NO_3^-, NH_4^+, $Si(OH)_4$, TOC, TON, and turbidity for all oceanside transects. Stations grouped by shore type showed that those stations situated along channels/passes possessed higher nutrient concentrations, phytoplankton biomass, and turbidity than those stations off land. These differences were very small but it is not known if they are biologically important. However, the fact that the benthic communities are different between these two habitats indicates that there may be some long term effects.

Probably the most interesting result of our data analysis was the elucidation of temporal trends in TP for much of the FKNMS. Trend analysis showed statistically significant increases in TP for the Tortugas, Marquesas, Lower Keys, and portions of the Middle and Upper Keys. These trends were remarkably linear and show little seasonality. The increases in TP were system wide and occurred outside the FKNMS on the SW Shelf as well. Rates of increase ranged from 0.01-0.07 μM yr^{-1} which was significant considering initial concentrations to be ~0.1-0.2 μM. No trends in TP were observed in Florida Bay or in those FKNMS sites most influenced by transport of Florida Bay waters. The effect of increased TP on the phytoplankton biomass has not been shown to be significant; i.e. no concurrent increases in Chl a were observed.

At this time we can only speculate as to the cause of these increases in TP concentrations but it is clear that the increases are driven by regional circulation patterns arising from the Loop and Florida Currents. We have begun the process of gathering information as to potential TP sources and transport mechanisms.

II. Project Background

The Florida Keys are a archipelago of sub-tropical islands of Pleistocene origin which extend in a NE to SW direction from Miami to Key West and out to the Dry Tortugas (Fig. 1). In 1990, President Bush signed into law the Florida Keys National Sanctuary and Protection Act (HR5909) which designated a boundary encompassing >2,800 square nautical miles of islands, coastal waters, and coral reef tract as the Florida Keys National Marine Sanctuary (FKNMS). The Comprehensive Management Plan (NOAA 1995) required the FKNMS to have a Water Quality Protection Plan (WQPP) thereafter developed by EPA and the State of Florida (EPA 1995). The contract for the water quality monitoring component of the WQPP was subsequently awarded to the Southeast Environmental Research Program at Florida International University and the field sampling program began in March 1995.

The waters of the FKNMS are characterized by complex water circulation patterns over both spatial and temporal scales with much of this variability due to seasonal influence in regional circulation regimes. The FKNMS is directly influenced by the Florida Current, the Gulf of Mexico Loop Current, inshore currents of the SW Florida Shelf (Shelf), discharge from the Everglades through the Shark River Slough, and by tidal exchange with both Florida Bay and Biscayne Bay. Advection from these external sources has significant effects on the physical, chemical, and biological composition of waters within the FKNMS, as may internal nutrient loading and freshwater runoff from the Keys themselves. Water quality of the FKNMS may be directly affected both by external nutrient transport and internal nutrient loading sources. Therefore, the geographical boundary of the FKNMS must not be thought of as enclosing a distinct ecosystem but rather as being one of political/regulatory definition.

Ongoing quarterly sampling of >200 stations in the FKNMS and Shelf, as well as monthly sampling of 100 stations in Florida Bay, Biscayne Bay, and the mangrove estuaries of the SW coast, has provided us with a unique opportunity to explore the spatial component of water quality variability. By stratifying the sampling stations according to depth, regional geography, distance from shore, proximity to tidal passes, and influence of Shelf waters we report some preliminary conclusions as to the relative importance of external vs. internal factors on the ambient water quality within the FKNMS.

III. Methods

Site Characteristics and Sampling Design

A spatial framework for FKNMS water quality management was proposed on the basis of geographical variation of regional circulation (Klein and Orlando, 1994). The final implementation plan (EPA, 1995) partitioned the FKNMS into 9 segments which was collapsed to 7 for routine sampling (Fig. 1). Station locations were developed using a stratified random design along onshore/offshore transects in Seg. 5, 7, and 9 or within EMAP grid cells in Seg. 1, 2, 4, and 6.

Segment 1 (Tortugas) includes the Dry Tortugas National Park and surrounding waters and is most influenced by the Loop Current and Dry Tortugas Gyre. Segment 2 (Marquesas) includes the Marquesas Keys and a shallow sandy area between the Marquesas and Tortugas called the Quicksands; Segment 4 (Backcountry) contains the shallow, hard-bottomed waters of the gulfside Lower Keys. Segments 2 and 4 are both influenced by water moving south from the Shelf. Segment 6 can be considered as part of western Florida Bay. This area is referred to as the Sluiceway as it heavily influenced by transport from Florida Bay and Shark River Slough (Smith, 1994). Segment 5 (Lower Keys), 7 (Middle Keys), and 9 (Upper Keys) include the inshore, Hawk Channel and reef tract of the Atlantic side of the Florida Keys. The Lower Keys are most influenced by cyclonic gyres spun off of the Florida Current, the Middle Keys by exchange with Florida Bay, while the Upper Keys are influenced by the Florida Current frontal eddies and to a certain extent by exchange with Biscayne Bay. All three oceanside segments are also influenced by wind and tidally driven lateral Hawk Channel transport (Pitts, 1997).

Spatial Analysis

Stations were grouped four different ways for statistical analysis: by surface or bottom samples, surface by segment, surface by transect distance, and surface by shore type. These groupings were subjectively defined using best available knowledge in an effort to provide information as to source, transport, and fate of water quality components. For the first grouping,

stations were selected as being >3 m depth where both surface and bottom samples were collected and stratified by depth. The second grouping included surface samples stratified by segment (Fig. 1) in accordance with the implementation plan (EPA, 1995). The third grouping consisted of those surface stations situated on ocean-side transects being aggregated according to their distance from shore: Alongshore, Hawk Channel, or Reef Tract. In addition, we initiated a transect of stations in the Tortugas off Loggerhead Key to serve as a reference. Since sampling at these locations in the Tortugas were only recently set up to address this question, the data is more sparse. Also there are only two "channel" stations in the Tortugas which makes the data more susceptible to outlier conditions.

One of the concerns of this program is to determine the contribution of water movement through the passes of the Keys to the water quality of the reef. To this end we decided to characterize the last grouping of transects as to shore type: those that are adjacent to land off Biscayne National Park off Old Rhodes Key, Elliot Key and the Safety Valve (BISC), those that abut land in Key Largo, Middle, and Lower Keys (LAND), and those transects which are aligned along an open channel or pass through the Keys (PASS). These grouping strategies may be changed when enough data is collected (ca. 5-7 yr) to be analyzed using a statistically objective, multivariate approach as has been done previously for Florida Bay and Ten Thousand Islands (Boyer et al., 1997; Boyer and Jones, 1998).

Typical water quality variables are usually skewed to the left resulting in non-normal distributions. Therefore it is more appropriate to use the median as the measure of central tendency. Data distributions of selected water quality variables are reported as box-and-whiskers plots. The box-and-whisker plot is a powerful statistic as it shows the median, range, the data distribution as well as serving as a graphical, nonparametric ANOVA. The center horizontal line of the box is the median of the data, the top and bottom of the box are the 25th and 75th percentiles (quartiles), and the ends of the whiskers are the 5th and 95th percentiles. The notch in the box is the 95% confidence interval of the median. When notches between boxes do not overlap, the medians are considered significantly different. Outliers ($<5^{th}$ and $>95^{th}$ percentiles) were excluded from the graphs to reduce visual compression. Differences in variables were also tested between groups using the Wilcoxon Ranked Sign test (comparable to t-test) and among groups by the Kruskall-Wallace test (ANOVA) with significance set at $P<0.05$.

In an effort to elucidate the contribution of external factors to the water quality of the FKNMS and to visualize gradients in water quality over the region, we combined data from other portions of our water quality monitoring network: Florida Bay, Biscayne Bay, Whitewater Bay, Ten Thousand Islands, and the Shelf (Fig. 1). Data from these 153 additional stations was collected during the same month as the FKNMS surveys and analyzed by the SERC laboratory using similar methodology and quality control as previously described.

Time Series Analysis

Data for the complete period of record were plotted as time series graphs (see separate Data Appendix) to illustrate any temporal trends that might have occurred. Trends were quantified by simple regression with significance set at P<0.10.

IV. Results and Discussion

Spatial Analysis

Summary statistics for all measured parameters split out by segment are shown in Table 2. This summary includes data from all sampling dates and stations for the period of record listed by median value (Median), minimum value (Min.), maximum value (Max.), and number of samples (*n*). Typical water quality data is skewed to the low end which results in a non-normal distribution, therefore, it is more appropriate to use the median as the measure of central tendency.

Distinguishing internal from external sources of nutrients in the FKNMS is a difficult task. The finer discrimination of internal sources into natural and anthropogenic inputs is even more difficult but most important as anthropogenic inputs may be regulated and possibly controlled by management activities. Advective transport of nutrients through the FKNMS was not measured by the existing fixed sampling plan. However, nutrient distribution patterns may be compared to the regional circulation regimes in an effort to visualize the contribution of external sources and advective transport to internal water quality of the FKNMS.

Circulation in coastal South Florida is dominated by regional currents such as the Loop Current, Florida Current, and Tortugas Gyre and by local transport via Hawk Channel and along-shore Shelf movements (Klein and Orlando, 1994). Regional currents may influence water quality over large areas by the advection of external surface water masses into and through the FKNMS (Lee et al., 1994) and by the intrusion of deep offshore ocean waters onto the reef tract as internal bores (Leichter et al., 1996). Local currents become more important in the mixing and transport of freshwater and nutrients from terrestrial sources (Smith, 1994; Pitts, 1997).

Spatial patterns of salinity in coastal South Florida show these major sources of freshwater to have more than just local impacts (see salinity figures in Appendix 1). In Biscayne Bay, freshwater is released through the canal system operated by SFWMD; the impact is clearly seen to affect northern Key Largo by causing a depression in median salinity coupled with high variability in alongshore sites. Freshwater entering NE Florida Bay via overland flow from Taylor Slough and C-111 basin in ENP can be seen to mix in a SW direction. The extent of influence of freshwater from Florida Bay on alongshore salinity in the Keys is less than that of Biscayne Bay but it is more episodic. Transport of low salinity water from Florida Bay does not affect the Middle Keys sites enough to depress the median salinity in this region but is manifested as increased variability. On the west coast, the large influence of the Shark River Slough, which drains the bulk of the Everglades and exits through the Whitewater Bay - Ten Thousand Islands mangrove complex, is clearly seen as impacting the Shelf waters. The mixing of Shelf waters with the Gulf of Mexico produces a salinity gradient in a SW direction which extends out to Key West. This freshwater source does not seem to impact the Backcountry because of its shallow nature but instead follows a trajectory of entering western Florida Bay and exiting out through the channels in the Middle Keys (Smith, 1994). This net transport of lower salinity water from mainland to reef in open channels through the Keys is observed more so as an increase in the range and variability of salinity than as a large depression in salinity.

In addition to surface currents there is evidence that internal tidal bores regularly impact the Key Largo reef tract (Leichter et al. 1996). Internal bores are episodes of higher density, deep water intrusion onto the shallower shelf or reef tract. Depending on their energy, internal tidal bores can promote stratification of the water column or cause complete vertical mixing as a breaking internal wave of sub-thermocline water. To determine the extent of stratification we

calculated the difference between surface and bottom density, delta sigma-t ($\Delta\sigma_t$), where positive values denoted greater density of bottom water relative to the surface. The resulting graph of $\Delta\sigma_t$ (Fig. 2), shows that the SW area of the Tortugas segment tends to experience the greatest frequency of stratification events. The decreased temperature and increased salinity in bottom waters from intrusion of deeper denser oceanic waters to this region may also account for increases in NO_3^-, TP, and SRP in these bottom waters as well. For example, in April 1998 a mass of colder, nutrient laden water from the Gulf of Mexico moved up onto the Tortugas reefs and fueled a large benthic macroalgae bloom (J. Porter, personal comm.). This event was observed throughout most of the eastern Gulf as far north as Pensacola. At the two most SW stations, temperatures dropped $\sim$4°C, NO_3^- increased 3 orders of magnitude, SRP and $Si(OH)_4$ increased by a factor of 100, while TP, turbidity, and in vivo Chl a specific fluorescence (measured via CTD) all doubled. As there was only a small increase in NH_4^+ during this event we believe the general case of elevated NH_4^+ and turbidity found in bottom waters throughout the FKNMS is most probably due to benthic flux and resuspension and not to subthermocline advection.

Surface $Si(OH)_4$ concentrations exhibited a pattern similar to salinity. The source of $Si(OH)_4$ in this geologic area of carbonate rock and sediments is from siliceous periphyton (diatoms) growing in the Shark River Slough, Taylor Slough, and C-111 basin watersheds. Unlike the Mississippi River plume with Chl a concentrations of 76 μg l^{-1} (Nelson and Dortch 1996), phytoplankton biomass on the Shelf (1-2 μg l^{-1} Chl a) was not sufficient to account for the depletion of $Si(OH)_4$ in this area. Therefore, $Si(OH)_4$ concentrations on the Shelf were rapidly depleted by mixing and by chemical precipitation (Moore et al., 1986) allowing $Si(OH)_4$ to be used as a semi-conservative tracer of freshwater in this system (Ryther et al. 1967). Unlike Florida Bay and the west coast, there was very little $Si(OH)_4$ loading to southern Biscayne Bay, mostly because the source of freshwater to this system is from canals which drain agricultural and urban areas of Dade County.

In the Lower and Middle Keys, it is clear that the source of $Si(OH)_4$ to the nearshore Atlantic waters is through the Sluiceway and Backcountry. $Si(OH)_4$ concentrations near the coast were elevated relative to the reef tract with much higher concentrations occurring in the Lower and Middle Keys than the Upper Keys (Fig. 5). There is an interesting peak in $Si(OH)_4$ concentration

in an area of the Sluiceway which is densely covered with the seagrass, *Syringodium* (Fourqurean et al., in press). We are unsure as to the source but postulate that it may be due to benthic flux.

Visualization of spatial patterns of NO_3^- concentration over South Florida waters provide an extended view of source gradients over the region (Appendix 1). Biscayne Bay, Florida Bay, and the Shark River area of the west coast exhibited high NO_3^- concentrations relative to the FKNMS and Shelf. Elevated NO_3^- in Biscayne Bay is the result of loading from both the canal drainage system and from inshore groundwater (Alleman et al., 1995; Meeder et al., 1997). The source of NO_3^- to Florida Bay is the Taylor Slough and C-111 basin (Boyer and Jones, 1999; Rudnick et al., 1999) while the Shark River Slough impacts the west coast mangrove rivers and out onto the Shelf. The oceanside transects off Biscayne Bay in Seg. 9 exhibited the lowest NO_3^- alongshore compared to the Middle and Lower Keys. A similar pattern was observed in a previous transect survey from these areas (Szmant and Forrester, 1996). They also showed an inshore elevation of NO_3^- relative to Hawk Channel and the reef tract which is also demonstrated in our analysis. Interestingly, NO_3^- concentrations in all stations in the Tortugas transect were similar to those of reef tract sites in the mainland Keys; there was no inshore elevation of NO_3^- on the transect off uninhabited Loggerhead Key.

A distinct intensification of NO_3^- occurs in the Backcountry region. Part of this increase may due to a local sources of NO_3^-, i.e. septic systems and stormwater runoff around Big Pine Key (Lapointe and Clark, 1992). However, there is another area, the Snipe Keys, that exhibits high NO_3^- which is uninhabited by man. This rules out the premise of septic systems being the only source of NO_3^- in this area. It is important to note that the Backcountry area is very shallow (~0.5 m) and hydraulically isolated from the Shelf and Atlantic which results in its having a relatively long water residence time. Elevated NO_3^- concentrations may be partially due to simple evaporative concentration as is seen in salinity.

Dissolved NH_4^+ concentrations were distributed in a similar manner as NO_3^- with highest concentrations occurring in Florida Bay, the Ten Thousand Islands, and the Backcountry (Appendix 1). NH_4^+ concentrations were very low in Biscayne Bay because it is not a major component of loading from the canal drainage system. NH_4^+ also showed similarities with NO_3^- in its spatial distribution, being lowest in the Upper Keys and highest inshore relative to offshore. There was no alongshore elevation of NH_4^+ concentrations in the Tortugas where levels were

similar to those of reef tract sites in the mainland Keys. That the least developed portion of the Upper Keys in Biscayne National Park and uninhabited Loggerhead Key (Tortugas) exhibited lowest NO_3^- and NH_4^+ concentrations is evidence of a local anthropogenic source for both of these variables along the ocean side of the Upper, Middle, and Lower Keys. This pattern of decline implies an onshore source which is diluted with distance from land by low nutrient Atlantic Ocean waters.

Elevated DIN concentrations in the Backcountry, on the other hand, are not so easily explained. We postulate that the high concentrations found there are due to a combination of anthropogenic loading, physical entrapment, and benthic N_2 fixation. The relative contribution of these potential sources is unknown. Lapointe and Matzie (1996) have shown that stormwater and septic systems are responsible for increased DIN loading in and around Big Pine Key. The effect of increased water residence time in DIN concentration is probably small. Salinities in this area were only 1-2 ppt higher than local seawater which resulted in a concentration effect of only 5-6%. Benthic N_2 fixation may potentially be very important in the N budget of the Backcountry. Measured rates of N_2 fixation in a *Thalassia* bed in Biscayne Bay, having very similar physical and chemical conditions, were 540 μmol N m^{-2} d^{-1} (Capone and Taylor, 1980). Without the plant community N demand, one day of N_2 fixation has the potential to generate a water column concentration of >1 μM NH_4^+ (0.5 m deep). Much of this NH_4^+ is probably nitrified and may help account for the elevated NO_3^- concentrations observed in this area as well. Clearly, N_2 fixation may be a significant component of the N budget in the Backcountry and that it may be a exported as DIN to the FKNMS in general.

Spatial patterns in TP in South Florida coastal waters were strongly driven by the west coast outputs (Appendix 1). A declining gradient in TP extended from the inshore waters of Whitewater Bay - Ten Thousand Islands mangrove complex out onto the Shelf and Tortugas. A declining gradient also extended from north central Florida Bay to the Middle Keys. Brand (1997) has postulated that groundwater from a subterranean Miocene quartz sand channel, "the river of sand", containing high levels of phosphorus is the source of TP in this region. However, no evidence of this source exists to date and the data from Florida Bay does not indicate a subterranean source either (Boyer and Jones unpublished data). A very small TP gradient was

seen NE Florida Bay signifying that Taylor Slough and the C-111 basin contribute little TP to the system. Finally, there was no evidence of a significant terrestrial source of TP to Biscayne Bay.

In the Keys, there was evidence of elevated TP in alongshore stations of the Middle and Lower Keys but the differences were very small. The Upper Keys actually showed higher TP concentrations on the reef tract than inshore implying an offshore source. Interestingly, the Tortugas area had higher TP concentrations than the Upper Keys as a result of Shelf water advection.

In South Florida coastal waters, very little of TP is found in the inorganic form (SRP - PO_4^{3-}); most is organic P (TOP). The distribution of SRP on the west coast and Shelf was similar to that of TP with the general gradient from the west coast to Tortugas remaining. However, the SRP distribution was distinctly different from that of TP in Florida Bay, Whitewater Bay, and Biscayne Bay. In central Florida Bay the N-S gradient previously observed for TP was highly diminished for SRP indicating that almost all the TP in central Florida Bay was in the form of TOP. It is unlikely that the source of TOP to this region is from overland flow or groundwater as this is also the region that expresses highest salinity. Alternately, we hypothesize that the presence of the Flamingo channel, running parallel to the southern coastline of Cape Sable, acts as a tidal conduit for episodic advection of inshore Shelf water to enter north central Florida Bay. Subsequent trapping and evaporation then may act to concentrate TOP in this region. The second difference in P distributions was that there was a significant SRP gradient present in NE Florida Bay that was not observed for TP. The sources of SRP to this area are the Taylor Slough and C-111 basin (W. Walker per. communication; Boyer and Jones, 1999; Rudnick et al., 1999).

Whitewater Bay displayed an east-west gradient in SRP concentrations which increased with salinity leading us to conclude that the freshwater inputs from the Everglades were not a source of SRP to this area. Finally, there was evidence of a significant onshore-offshore SRP gradient in southern Biscayne Bay; most probably as a direct result of canal loading and groundwater seepage to this region (Meeder et al., 1997).

Concentrations of TOC (Appendix 1) and TON (not shown) were remarkably similar in pattern of distribution across the South Florida coastal hydroscape. The decreasing gradient from west coast to Tortugas was very similar to that of TP. A steep gradient with distance from land was observed in Biscayne Bay. Both these gradients were due to terrestrial loading. On the west

coast, the source of TOC and TON was from the mangrove forests. Our data from this area shows that concentrations of TOC and TON increased from Everglades headwaters through the mangrove zone and then decrease with distance offshore. In Biscayne Bay, much of the TOC and TON is from agricultural land use. The high concentrations of TOC and TON found in Florida Bay were due to a combination of terrestrial loading (Boyer and Jones, in press), *in situ* production by seagrass and phytoplankton, and evaporative concentration (Fourqurean et al., 1993).

Advection of Shelf and Florida Bay waters through the Sluiceway and passes accounted for this region and the inshore area of the Middle Keys as having highest TOC and TON of the FKNMS. Strong offshore gradients in TOC and TON existed for all mainland Keys segments but not for the Tortugas transect. Part of this difference may be explained by the absence of mangroves in the single Tortugas transect. The higher concentrations of TOC and TON in the inshore waters of the Keys then implies a terrestrial source rather than simply benthic production and sediment resuspension. Main Keys reef tract concentrations of TOC and TON were similar to those found in the Tortugas.

Much emphasis has been placed on assessing the impact of episodic phytoplankton blooms in Florida Bay on the offshore reef tract environment. Spatial patterns of Chl *a* concentrations (Appendix 1) showed that NW Florida Bay, Whitewater Bay, and the Ten Thousand Islands exhibited high levels of Chl *a* relative to Biscayne Bay, Shelf, and FKNMS. The highest Chl *a* concentrations were found in west coast mangrove estuaries (up to 45 μg l^{-1} in Alligator Bay, TTI). Chl *a* is also routinely high (~2 μg l^{-1}) in NW Florida Bay along the channel connecting the Shelf to Flamingo, ENP. It is interesting that Chl *a* concentrations are higher in the Marquesas (0.36 μg l^{-1}) than in other areas of the FKNMS. When examined in context with the whole South Florida ecosystem, it is obvious that the Marquesas zone should be considered a continuum of the Shelf rather than a separate management entity. This shallow sandy area (often called the Quicksands) acts as a physical mixing zone between the Shelf and the Atlantic Ocean and is a highly productive area for other biota as well as it encompasses the historical Tortugas shrimping grounds. A Chl *a* concentration of 1 μg l^{-1} in the water column of a reef tract is considered a problem as it indicates potential of eutrophication. On the other hand, a similar Chl *a* level in the Quicksands indicates a productive shrimp fishery.

The oceanside transects in the Upper Keys (Seg. 9) exhibited the lowest overall Chl *a* concentrations of any zone in the FKNMS. Ocean transects showed a slight increase in Chl *a* on the reef tract in this area. Transects off the Middle and Lower Keys showed that a drop in Chl *a* occurred only in the reef tract sites; there was no linear decline with distance from shore (data not shown). alongshore compared to the Middle and Lower Keys. Interestingly, Chl *a* concentrations in the Tortugas transect showed a similar pattern as the mainland Keys. Inshore and Hawk Channel Chl *a* concentrations among Middle Keys, Lower Keys and Tortugas sites were not significantly different. As inshore Chl *a* concentrations in the Tortugas were similar to those in the Middle and Lower Keys, we see no evidence of phytoplankton bloom transport from Florida Bay under this type of sampling design. There was however some slight evidence of increased Chl *a* in those stations along the major passes in the Keys relative to those abutting land. The differences between these two groupings were very small (0.25 vs. 0.20 μg l^{-1}).

Along with P concentration, turbidity is probably the second most important determinant of local ecosystem health. The fine, low density carbonate sediments in this area are easily resuspended, rapidly transported, and have high light scattering potential per gram of material. High water column turbidity and transport directly affects filter feeding organisms by clogging their feeding apparatus and by increasing local sedimentation rate. Sustained high turbidity of the water column indirectly affects benthic community structure by decreasing light penetration, promoting seagrasses extinction. Large scale observations of turbidity clearly show patterns of onshore-offshore gradients which extend out onto the Shelf to the Marquesas (Appendix 1). In the last seven years, turbidities in Florida Bay have increased dramatically in the NE and central regions (Boyer et al. 1998) potentially as a consequence of destabilization of the sediment from seagrass die-off (Robblee et al., 1991).

Strong turbidity gradients were observed for all Keys transects but reef tract levels were remarkably similar regardless of inshore levels. High alongshore turbidity is most probably due to the shallow water column being easily resuspended by wind and wave action. Inshore stations in the Middle Keys had higher turbidity than other segments. Transects aligned with major passes had slightly greater turbidity than those against land but the difference was not statistically significant. Light extinction (K_d) was highest alongshore and improved with distance from land

(data not shown). This trend was expected as light extinction is directly related to the turbidity of the water.

Using the DIN:TP ratio is a relatively simple method of determining phytoplankton nutrient limitation status of the water column (Redfield, 1967). Most of the FKNMS was shown to have DIN:SRP values < 16:1, indicating the potential for phytoplankton to be limited by N at these sites. The bulk of Florida Bay and both southern and northern Biscayne Bay were severely P limited, mostly as a result of high DIN concentrations. The south-north shift from P to N limitation observed in the west coast estuaries has been ascribed to changes in landuse and bedrock geochemistry of the watersheds (Boyer and Jones, 1998). The west coast south of 25.4 N latitude is influenced by overland freshwater flow from the Everglades and Shark River Slough having very low P concentrations relative to N. Above 25.7 N latitude the bedrock geology of the watershed changes from carbonate to silicate based and landuse changes from relatively undeveloped wetland (Big Cypress Basin) to a highly urban/agricultural mix (Naples, FL).

Time Series Analysis

We did not expect to see any temporal trends in the data because of the short data record (only 17 points on the graph), the usually high variability of the data, and the potential interference of a poorly resolved seasonal signal. This was true for all measured variables except TP. Trend analysis showed statistically significant increases in TP for the Tortugas, Marquesas, Lower Keys, and portions of the Middle and Upper Keys (Fig. 3). These trends were remarkably linear and showed little seasonality. Rates of increase ranged from 0.01-0.07 µM yr^{-1} which was especially significant considering initial concentrations to be ~0.1-0.2 µM (Table 2). The effect of increased TP on the phytoplankton biomass was not shown to be significant; i.e. no concurrent increases in Chl a were observed.

The trend in TP was system wide and occurred outside the FKNMS on the SW Shelf as well (Fig. 4). It is important to emphasize that this trend was a regional phenomena and was not due to local inputs from the Florida Keys alone. These increases must also be put in perspective with other ecological changes occurring in the region. No trends in TP were observed in the western Florida Bay/Inner Shelf zone or in those FKNMS sites most influenced by transport of Florida Bay waters. During the same time period as this study, TP concentrations in Florida Bay proper

were declining (Boyer et al. 1999). The absence of TP trends in the Middle and Lower Keys may have been due to the influence of low TP Florida Bay waters. This is a marked departure from the more alarmist thinking of some scientists and public that Florida Bay is source of nutrients to the reef and a major cause of reef decline.

A simple model of potential forcing functions on TP concentration (or any other non-conservative species) is shown in Fig. 5. The local concentration of a biologically reactive nutrient such as TP may be affected by advective transport into or out of the compartment (tides, currents, etc.), benthic flux into or out of the water column (including sedimentation), biological uptake and remineralization in the water column, and atmospheric input.

At this time we can only speculate as to the cause of these increases in TP concentrations but it is clear that much of the trend is driven by regional circulation patterns arising from the Loop Current which entrains water from other coastal estuaries such as the Caloosahatchee River and Tampa Bay as well as the Mississippi. That the increases have occurred in deep and shallow water stations at both the surface and bottom over a consistent period of time rules out episodic upwelling as a major factor.

We know of no data which addressed changes in internal cycling processes (benthic flux or water column cycling) over this period. However, there is some preliminary evidence that seagrass may be responsible for some of the trend patterns. Areas where TP trends were absent were also described as areas of dense seagrass beds (Fourqurean personal comm.). One hypothesis is that the potential increase in TP concentration in these areas was modulated by uptake by the seagrass community and therefore showed no significant change.

Summary

The large scale of this monitoring program has allowed us to assemble a much more holistic view of broad physical/chemical/biological interactions occurring over the South Florida hydroscape. Much information has been gained by inference from this type of data collection program: major nutrient sources have be confirmed, relative differences in geographical determinants of water quality have been demonstrated, large scale transport via circulation pathways have been elucidated and temporal trends are becoming evident. In addition we have shown the importance of looking "outside the box" for questions asked within. Rather than

thinking of water quality monitoring as being a static, non-scientific pursuit it should be viewed as a tool for answering management questions and developing new scientific hypotheses.

References

ALLEMAN, R. W., ET AL. 1995. Biscayne Bay surface water improvement and management. Technical supporting document. South Florida Water Management District.

BOYER, J. N., J. W. FOURQUREAN, AND R. D. JONES. 1997. Spatial characterization of water quality in Florida Bay and Whitewater Bay by multivariate analysis: Zones of similar influence (ZSI). Estuaries **20**: 743-758.

BOYER, J. N., J. W. FOURQUREAN, AND R. D. JONES. 1999. Seasonal and long term trends in the water quality of Florida Bay (1989 - 1997). Estuaries **22**: 417-430.

BOYER, J. N., AND R. D. JONES. 1998. Influence of coastal morphology and watershed characteristics on the water quality of mangrove estuaries in the Ten Thousand Islands - Whitewater Bay complex. Proceedings of the 1998 Florida Bay Science Conference. University of Florida Sea Grant.

BOYER, J. N., AND R. D. JONES. 1999. Effects of freshwater inputs and loading of phosphorus and nitrogen on the water quality of Eastern Florida Bay, p. 545-561. *In* K. R. Reddy, G. A. O'Connor, and C. L. Schelske (eds.) Phosphorus biogeochemistry in sub-tropical ecosystems. CRC/Lewis Publishers, Boca Raton, Florida.

BRAND, L. 1998. The role of groundwater in the Florida Bay ecosystem. Proceedings of the 1998 Florida Bay Science Conference. University of Florida Sea Grant.

CAPONE, D. G., AND B. F. TAYLOR. 1980. Microbial nitrogen cycling in a seagrass community, p. 153-161. *In* V. S. Kennedy [ed.], Estuarine perspectives. Academic.

ENVIRONMENTAL PROTECTION AGENCY. 1995. Water quality protection program for the Florida Keys National Marine Sanctuary: Phase III report. Final report submitted to the Environmental Protection Agency under Work Assignment 1, Contract No. 68-C2-0134. Battelle Ocean Sciences, Duxbury, MA and Continental Shelf Associates, Inc., Jupiter FL.

FOURQUREAN, J.W., M.D. DURAKO, M.O. HALL AND L.N. HEFTY. (in press). Seagrass distribution in south Florida: a multi-agency coordinated monitoring program. *In* Porter, J.W. and K.G. Porter, eds. Linkages between ecosystems in the south Florida hydroscape: the river of grass continues.

KLEIN, C. J., AND S. P. ORLANDO JR. 1994. A spatial framework for water-quality management

in the Florida Keys National Marine Sanctuary. Bull. Mar. Sci. **54**: 1036-1044.

LAPOINTE, B. E., AND M. W. CLARK. 1992. Nutrient inputs from the watershed and coastal eutrophication in the Florida Keys. Estuaries **15**: 465-476.

LAPOINTE, B. E., AND W. R. MATZIE. 1996. Effects of stormwater nutrient discharges on eutrophication processes in nearshore waters of the Florida Keys. Estuaries **19**: 422-435.

LEE, T. N., M. E. CLARKE, E. WILLIAMS, A. F. SZMANT, AND T. BERGER. 1994. Evolution of the Tortugas gyre and its influence on recruitment in the Florida Keys. Bull. Mar. Sci. **54**: 621-646.

LEICHTER, J. J., S. R. WING, S. L. MILLER, AND M. W. DENNY. 1996. Pulsed delivery of subthermocline water to Conch Reef (Florida Keys) by internal tidal bores. Limnol. Oceanogr. **41**: 1490-1501.

MEEDER, J. F., J. ALVORD, M. BYRNS, M. S. ROSS, AND A. RENSHAW. 1997. Distribution of benthic nearshore communities and their relationship to groundwater nutrient loading. Final report to Biscayne National Park.

MOORE, W. S., J. L. SARMIENTO, AND R. M. KEY. 1986. Tracing the Amazon component of surface Atlantic water using ^{228}Ra, salinity, and silica. J. Geophys. Res. **91:** 2574-2580.

NELSON, D. M., AND Q. DORTCH. 1996. Silicic acid depletion and silicon limitation in the plume of the Mississippi River: evidence from kinetic studies in spring and summer. Mar. Ecol. Prog. Ser. **136:** 163-178.

NATIONAL OCEANIC AND ATMOSPHERIC ADMINISTRATION. 1995. Florida Keys National Marine Sanctuary Draft Management Plan/Environmental Impact Statement.

PITTS, P. A. 1997. An investigation of tidal and nontidal current patterns in Western Hawk Channel, Florida Keys. Cont. Shelf Res. **17**: 1679-1687.

ROBBLEE, M. B., T. B. BARBER, P. R. CARLSON JR., M. J. DURAKO, J. W. FOURQUREAN, L. M. MUEHLSTEIN, D. PORTER, L. A. YABRO, R. T. ZIEMAN, AND J. C. ZIEMAN. 1991. Mass mortality of the tropical seagrass *Thallassia testudinum* in Florida Bay (USA). Mar. Ecol. Prog. Ser. **71**: 297-299.

RUDNICK, D., Z. CHEN, D. CHILDERS, T. FONTAINE, AND J. N. BOYER. 1999. Phosphorus and nitrogen inputs to Florida Bay: the importance of the Everglades watershed. Estuaries

22: .

RYTHER, J. H., D. W. MENZE, AND N. CORWIN. 1967. Influence of the Amazon River outflow on the ecology of the western tropical Atlantic, I. Hydrography and nutrient chemistry. J. Mar. Res. **25**: 69-83.

SMITH, N. P. 1994. Long-term Gulf-to-Atlantic transport through tidal channels in the Florida Keys. Bull. Mar. Sci. **54**: 602-609.

SZMANT, A. M., AND A. FORRESTER. 1996. Water column and sediment nitrogen and phosphorus distribution patterns in the Florida Keys, USA. Coral Reefs **15**: 21-41.

List of Tables

Table 1. Summary statistics for each water quality variable in the FKNMS. Data are summarized as median (Median), minimum value (Min.), maximum value (Max.), and number of samples (*n*).

Table 2. Time series regression of total phosphorus. Statistically significant slopes (P<0.10) are shown in boldface.

Table 1.

Variable	Segment	Median	Minimum	Maximum	*n*
Surface	1	0.015	0.000	0.713	239
NO_3^-	2	0.022	0.000	1.153	303
(μM)	4	0.141	0.000	4.418	368
	5	0.090	0.000	1.550	452
	6	0.055	0.000	2.940	235
	7	0.066	0.000	2.094	278
	9	0.058	0.000	1.658	506
Bottom	1	0.072	0.000	4.455	239
NO_3^-	2	0.033	0.000	1.455	303
(μM)	4				
	5	0.091	0.000	1.483	356
	6	0.073	0.007	0.138	2
	7	0.045	0.000	2.305	202
	9	0.060	0.000	0.627	350
Surface	1	0.028	0.000	0.325	239
NO_2^-	2	0.035	0.000	0.223	303
(μM)	4	0.065	0.007	0.347	368
	5	0.050	0.003	0.273	457
	6	0.063	0.007	0.255	235
	7	0.050	0.000	0.258	282
	9	0.040	0.000	0.190	506
Bottom	1	0.040	0.003	1.732	240
NO_2^-	2	0.035	0.000	0.613	303
(μM)	4				
	5	0.045	0.000	0.200	360
	6	0.033	0.033	0.033	2
	7	0.043	0.000	0.688	205
	9	0.037	0.000	0.167	347
Surface	1	0.297	0.022	1.893	239
NH_4^+	2	0.324	0.025	1.703	302
(μM)	4	0.504	0.130	9.097	368
	5	0.333	0.000	2.160	457
	6	0.355	0.000	10.320	235
	7	0.390	0.005	2.442	282
	9	0.300	0.000	2.450	506

Bottom	1	0.287	0.007	1.320	240
NH_4^+	2	0.302	0.028	1.255	303
(μM)	4				
	5	0.334	0.000	2.377	360
	6	0.330	0.253	0.407	2
	7	0.307	0.000	1.885	205
	9	0.295	0.022	1.880	347
Surface	1	7.730	3.638	24.118	239
TON	2	8.859	3.809	24.928	302
(μM)	4	15.054	8.628	33.054	368
	5	11.097	3.557	27.278	452
	6	15.236	6.039	32.100	235
	7	11.470	4.693	42.682	276
	9	8.925	3.702	24.855	505
Bottom	1	7.163	2.254	18.825	238
TON	2	8.438	4.444	16.537	303
(μM)	4				
	5	9.742	4.013	22.927	355
	6	11.968	7.673	16.264	2
	7	9.755	3.788	26.637	201
	9	8.340	3.900	28.200	346
Surface	1	0.185	0.083	0.350	240
TP	2	0.205	0.049	0.660	303
(μM)	4	0.230	0.092	0.621	368
	5	0.175	0.000	0.382	457
	6	0.228	0.119	0.843	235
	7	0.170	0.010	0.576	282
	9	0.153	0.000	0.295	503
Bottom	1	0.192	0.085	0.638	239
TP	2	0.208	0.033	0.518	303
(μM)	4				
	5	0.173	0.000	0.395	358
	6	0.240	0.135	0.345	2
	7	0.169	0.010	0.391	204
	9	0.157	0.000	0.350	346
Surface	1	0.007	0.000	0.087	236
SRP	2	0.010	0.000	0.092	303
(μM)	4	0.010	0.000	0.297	368
	5	0.005	0.000	0.203	457
	6	0.007	0.000	0.258	235
	7	0.007	0.000	0.098	282
	9	0.005	0.000	0.120	506

Bottom SRP (μM)	1	0.010	0.000	0.390	238
	2	0.007	0.000	0.193	303
	4				
	5	0.007	0.000	0.195	360
	6	0.006	0.003	0.010	2
	7	0.007	0.000	0.092	206
	9	0.007	0.000	0.090	347
Surface APA (μM hr^{-1})	1	0.031	0.010	0.195	180
	2	0.038	0.013	0.840	230
	4	0.086	0.007	1.286	365
	5	0.050	0.014	0.370	454
	6	0.080	0.010	0.552	235
	7	0.062	0.010	0.434	273
	9	0.056	0.009	0.450	483
Bottom APA (μM hr^{-1})	1	0.029	0.010	0.080	179
	2	0.039	0.008	0.261	230
	4				
	5	0.046	0.000	0.428	358
	6	0.020	0.000	0.040	3
	7	0.053	0.000	0.491	198
	9	0.050	0.000	0.206	331
Surface Chl a (μg l^{-1})	1	0.232	0.000	1.347	236
	2	0.373	0.011	6.810	303
	4	0.289	0.000	6.388	368
	5	0.271	0.000	1.676	455
	6	0.247	0.067	1.970	234
	7	0.237	0.000	1.792	282
	9	0.202	0.000	2.698	506
Surface TOC (μM)	1	178.573	86.979	1054.792	238
	2	200.194	88.480	501.750	302
	4	242.552	136.000	1653.542	368
	5	203.560	93.438	674.042	453
	6	271.104	122.170	970.167	235
	7	211.688	98.083	805.310	281
	9	193.896	92.646	512.479	505
Bottom TOC (μM)	1	169.625	89.375	883.104	238
	2	195.521	94.940	847.708	303
	4				
	5	191.806	92.771	332.896	357
	6	286.386	126.730	446.042	2
	7	197.583	102.396	760.770	204
	9	187.792	92.833	482.500	346

Surface SI(OH)$_4$ (μM)	1	0.310	0.000	3.902	209
	2	0.537	0.000	4.992	265
	4	1.680	0.000	20.015	319
	5	1.337	0.000	16.035	397
	6	5.672	0.078	127.110	206
	7	1.027	0.000	37.362	246
	9	0.307	0.000	12.990	446
Bottom SI(OH)$_4$ (μM)	1	0.555	0.000	5.777	209
	2	0.675	0.000	6.923	265
	4				
	5	1.053	0.000	11.110	311
	6	0.273	0.273	0.273	1
	7	0.590	0.000	30.195	181
	9	0.284	0.000	11.360	306
Surface Turbidity (NTU)	1	0.295	0.000	3.000	223
	2	0.853	0.000	18.800	285
	4	1.010	0.000	11.345	357
	5	0.465	0.000	4.885	455
	6	0.825	0.000	37.000	234
	7	0.458	0.000	17.350	282
	9	0.345	0.000	8.800	505
Bottom Turbidity (NTU)	1	0.385	0.000	2.626	223
	2	1.100	0.000	9.100	285
	4	0.650	0.150	1.477	13
	5	0.475	0.000	4.885	366
	6	1.012	0.095	7.295	17
	7	0.270	0.000	16.900	211
	9	0.280	0.000	7.950	356
Surface Salinity (ppt)	1	36.200	32.300	36.600	237
	2	36.200	33.600	37.000	297
	4	36.000	30.500	38.600	368
	5	36.200	33.600	38.600	436
	6	35.800	29.900	40.300	234
	7	36.100	33.100	38.400	264
	9	36.100	22.244	37.800	493
Bottom Salinity (ppt)	1	36.200	34.000	37.000	238
	2	36.200	34.600	37.400	297
	4	36.000	33.400	38.700	365
	5	36.208	33.400	38.600	433
	6	35.800	29.900	39.700	232
	7	36.130	33.000	38.900	266
	9	36.100	21.738	37.800	471

Surface Temperature (°C)	1	26.100	21.100	31.100	239
	2	26.500	19.287	32.300	298
	4	27.900	19.100	36.100	368
	5	27.400	20.300	33.600	439
	6	28.500	20.900	32.700	234
	7	26.150	19.500	39.600	264
	9	25.900	17.300	32.200	494
Bottom Temperature (°C)	1	24.700	18.200	30.600	239
	2	25.695	19.000	32.200	298
	4	28.000	18.700	36.800	365
	5	27.000	20.100	33.400	435
	6	28.500	20.300	32.700	232
	7	25.800	19.400	32.900	267
	9	25.621	17.100	32.100	473
K_d (m^{-1})	1	0.126	0.037	0.700	240
	2	0.208	0.030	0.961	299
	4	0.326	0.026	2.562	326
	5	0.180	0.005	1.090	434
	6	0.370	0.021	1.389	177
	7	0.188	0.009	1.546	269
	9	0.203	0.007	1.573	443
Surface DIN:TP	1	2.189	0.000	11.026	239
	2	2.077	0.211	11.563	302
	4	3.215	0.584	32.306	368
	5	2.980	0.261	61.250	447
	6	2.203	0.333	35.390	235
	7	3.275	0.325	85.500	278
	9	2.723	0.149	71.250	501
Surface DO$_{sat}$ (%)	1	92.495	0.000	150.029	234
	2	92.623	0.000	114.926	294
	4	93.896	43.526	169.865	368
	5	91.783	46.935	153.343	431
	6	94.867	65.021	148.204	234
	7	92.553	68.049	124.391	262
	9	92.989	38.130	126.051	483
Bottom DO$_{sat}$ (%)	1	91.678	0.000	107.570	224
	2	92.546	0.000	113.452	295
	4	94.194	43.526	171.438	365
	5	91.969	46.935	143.602	427
	6	95.310	62.684	149.616	232
	7	93.061	46.915	127.425	261
	9	93.817	0.000	128.710	463

$\Delta\sigma_t$	1	0.197	-0.190	5.552	236
	2	0.059	-0.188	27.173	297
	4	0.000	-4.424	6.528	365
	5	0.076	-0.383	3.343	430
	6	0.000	-0.370	3.590	232
	7	0.070	-1.440	4.762	263
	9	0.031	-3.185	1.522	466

Table 2.

Station	Site	Slope	P
200	**Fowey Rocks**	**0.033**	**0.030**
201	**Sands Key**	**0.017**	**0.016**
202	Bowles Bank	0.009	0.430
203	Triumph Reef	0.017	0.118
204	Elliott Key	0.012	0.270
205	Margo Fish Shoal	0.011	0.330
206	**Ajax Reef**	**0.022**	**0.018**
207	**Old Rhodes Key**	**0.017**	**0.059**
208	Old Rhodes Key Channel	0.014	0.187
209	**Channel Key**	**0.028**	**0.004**
210	**Old Rhodes Key Reef**	**0.028**	**0.006**
211	Pennikamp G27	0.013	0.313
212	Turtle Harbor	0.011	0.273
213	**Turtle Reef**	**0.020**	**0.045**
214	Port Elizabeth	0.003	0.720
215	Carysfort Channel	0.007	0.450
216	Carysfort Reef	0.016	0.125
217	Rattlesnake Key	0.012	0.120
218	**White Bank**	**0.021**	**0.020**
219	The Elbow	0.009	0.440
220	Radabob Key	0.024	0.374
221	**Radabob Key Channel**	**0.019**	**0.067**
222	**Dixie Shoal**	**0.014**	**0.091**
223	**Mosquito Bank**	**0.017**	**0.057**
224	**Molasses Reef Channel**	**0.023**	**0.025**
225	**Molasses Reef**	**0.023**	**0.011**
226	**Tavernier Harbor**	**0.025**	**0.004**
227	**Triangles**	**0.034**	**0.001**
228	**Conch Reef**	**0.029**	**0.001**
229	**Plantation Point**	**0.035**	**0.001**
230	**The Rocks**	**0.027**	**0.006**
231	**Davis Reef**	**0.031**	**0.001**
232	**Upper Matecumbe Key**	**0.026**	**0.007**
233	**Upper Matecumbe Chnl**	**0.054**	**0.013**
234	**Fish Haven**	**0.029**	**0.004**
235	**Indian Key**	**0.036**	**0.005**
236	**Indian Key Channel**	**0.034**	**0.000**
237	**Indian Key Offshore**	**0.016**	**0.091**
238	**Matecumbe Harbor**	**0.025**	**0.018**
239	**Lower Matecumbe Chnl**	**0.025**	**0.019**

240	**Matecumbe Offshore**	**0.034**	**0.003**
241	**Long Key**	**0.029**	**0.013**
242	Long Key Channel	0.025	0.045
243	**Tennessee Reef**	**0.039**	**0.001**
244	**Long Key Pass Inshore**	**0.042**	**0.000**
245	**Long Key Pass Channel**	**0.029**	**0.001**
246	**Long Key Pass Offshore**	**0.022**	**0.004**
247	**Key Colony Beach**	**0.026**	**0.021**
248	Coffins Patch Channel	0.007	0.253
249	Coffins Patch Offshore	0.014	0.105
250	**Seven Mile Bridge**	**0.028**	**0.062**
251	Seven Mile Br. Channel	0.014	0.103
252	Seven Mile Br. Offshore	0.018	0.123
253	Spanish Harbor Keys	0.012	0.433
254	Bahia Honda Key	0.008	0.447
255	Bahia Honda Channel	0.006	0.547
256	**Bahia Honda Offshore**	**0.015**	**0.051**
257	**Long Beach**	**0.018**	**0.055**
258	**Big Pine Channel**	**0.015**	**0.078**
259	**Big Pine Shoal**	**0.016**	**0.081**
260	Newfound Harbor Keys	0.015	0.143
261	American Shoal Channel	0.013	0.139
262	**Looe Key Channel**	**0.017**	**0.025**
263	**Looe Key**	**0.024**	**0.015**
264	**Aquarius**	**0.026**	**0.025**
265		**0.020**	**0.059**
266	Tarpon Creek	0.013	0.221
267	**American Shoal**	**0.014**	**0.032**
268	**Saddlebunch Keys**	**0.026**	**0.026**
269	**West Washerwoman**	**0.023**	**0.060**
270	Maryland Shoal	0.016	0.118
271	**Boca Chica Key**	**0.028**	**0.041**
272	**Eastern Sambo**	**0.025**	**0.019**
273	**Eastern Sambo Offshore**	**0.030**	**0.000**
274	**Boca Chica Channel**	**0.053**	**0.001**
275	**Boca Chica Mid**	**0.041**	**0.004**
276	**Boca Chica Offshore**	**0.034**	**0.021**
277	**Key West Cut A**	**0.028**	**0.024**
278	**Western Head**	**0.026**	**0.066**
279	**Main Ship Channel**	**0.032**	**0.009**
280	**Eastern Dry Rocks**	**0.033**	**0.021**
281	**Middle Ground**	**0.040**	**0.002**

282	Arsenic Bank	0.006	0.710
283		-0.027	0.423
284	Tripod Bank	-0.010	0.489
285	**Channel Key Pass**	**0.034**	**0.023**
286	**Toms Harbor Cut**	**0.032**	**0.046**
287	**Bamboo Banks**	**0.032**	**0.037**
288		0.016	0.321
289	**Bamboo Key**	**0.036**	**0.035**
290	Bluefish Bank	-0.007	0.790
291	Bullard Bank	0.012	0.456
292	John Sawyer Bank	0.021	0.144
293	**Bethel Bank**	**0.024**	**0.075**
294	Red Bay Bank	0.015	0.212
295	Bullfrog Banks	0.022	0.395
296	W. Bahia Honda Key	0.012	0.393
297	Cocoanut Key	0.014	0.378
298	Harbor Key Bank	0.001	0.949
299	Bogie Channel	0.013	0.145
300	Little Pine Key	0.010	0.356
301	Cutoe Key	0.004	0.690
302	Content Passage	0.014	0.310
303	**Pine Channel**	**0.022**	**0.005**
304	**Toptree Hammock Chan.**	**0.030**	**0.006**
305	Cudjoe Key	0.016	0.284
306	Johnson Key Channel	0.016	0.215
307	Tarpon Belly Keys	0.010	0.314
308	**Kemp Channel**	**0.019**	**0.070**
309	Snipe Point	0.015	0.241
310	**Snipe Keys**	**0.030**	**0.026**
311	Shark Key	0.008	0.370
312	E. Harbor Key Channel	0.010	0.512
313	Lower Harbor Keys	0.022	0.110
314	**Howe Key Channel**	**0.037**	**0.003**
315	**Calda Channel**	**0.026**	**0.026**
316	**Man of War Harbor**	**0.026**	**0.000**
317	**Garrison Bight**	**0.025**	**0.002**
318	**KY Northwest Channel**	**0.021**	**0.017**
319	N Boca Grande Channel	0.021	0.137
320	Loggerhead Marker	0.018	0.163
321	**Loggerhead Channel**	**0.021**	**0.045**
322	**Satan Shoal**	**0.030**	**0.007**
323		**0.027**	**0.023**
324	**Ellis Rock**	**0.024**	**0.054**

325	SE Marquesas	0.043	0.140
326		**0.045**	**0.034**
327	**N Quicksands**	**0.031**	**0.010**
328	**Marquesas Rock**	**0.032**	**0.000**
329		**0.035**	**0.004**
330	**New Ground**	**0.036**	**0.009**
331		**0.039**	**0.026**
332	**S Quicksands**	**0.034**	**0.031**
333	**Half Moon Shoal**	**0.038**	**0.009**
334		**0.022**	**0.037**
335		**0.028**	**0.032**
336		**0.037**	**0.001**
337	**Rebecca Shoal**	**0.034**	**0.000**
338	**Garden Key**	**0.038**	**0.000**
339		**0.045**	**0.005**
340		**0.041**	**0.000**
341	**Northwest Channel**	**0.043**	**0.000**
342	**NE DTNP**	**0.041**	**0.000**
343	**N DTNP**	**0.041**	**0.000**
344	**Southwest Channel**	**0.036**	**0.001**
345		**0.039**	**0.001**
346	**W DTNP**	**0.045**	**0.000**
347	**Loggerhead Offshore**	**0.037**	**0.001**
348	**Hospital Key**	**0.035**	**0.002**
349	**Logerhead Inshore**	**0.037**	**0.000**
350		**0.032**	**0.000**

List of Figures

Fig. 1. The SERC Water Quality Monitoring Network showing the FKNMS boundary, segments, and distribution of sampling stations within the FKNMS, Florida Bay, Biscayne Bay, Whitewater Bay, Ten Thousand Islands, and Southwest Florida Shelf.

Fig. 2. Contour plot of median difference in density between surface and bottom waters ($\Delta\sigma_t$) for the period of record.

Fig. 3. Representative example of time series plot of TP (μM) with time at station #350 in the far SW corner of theTortugas segment. Note the consistent increase and absence of seasonal variation.

Fig. 4. Contour map of slope of trend line for significant regressions of TP (μM) with time.

Fig. 1
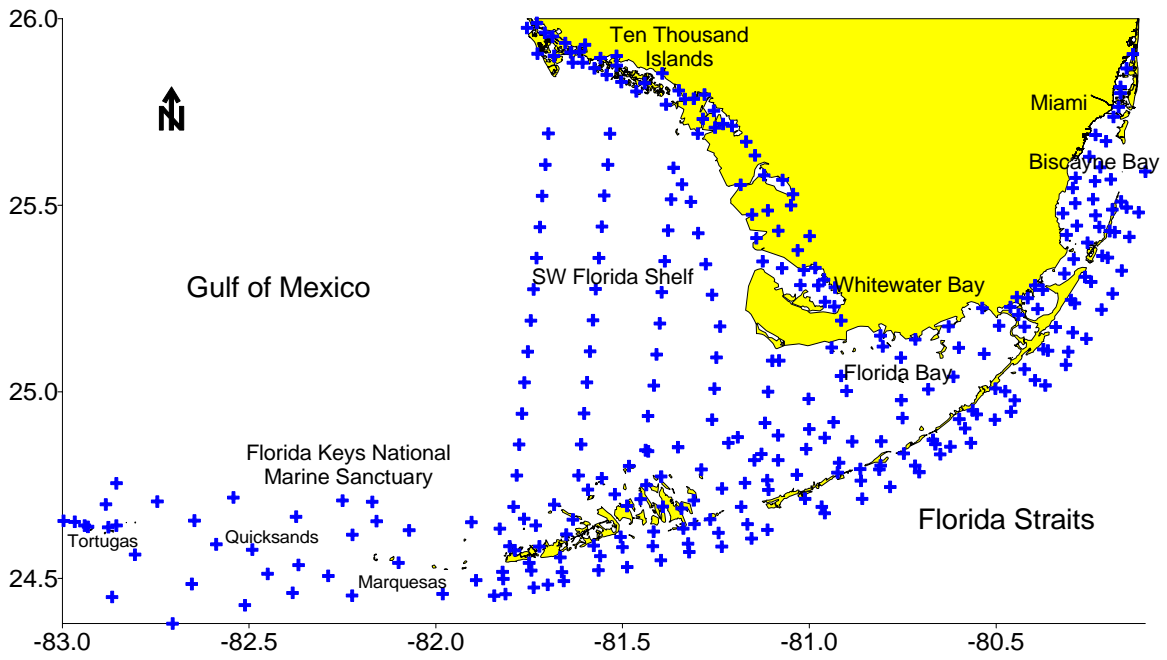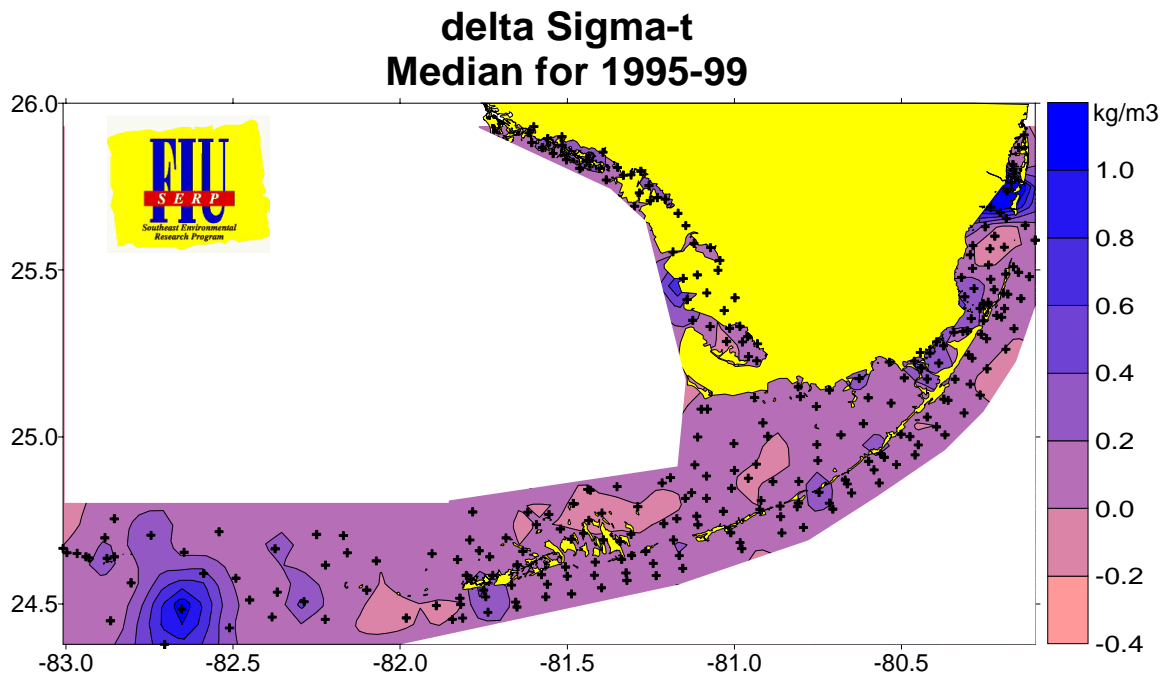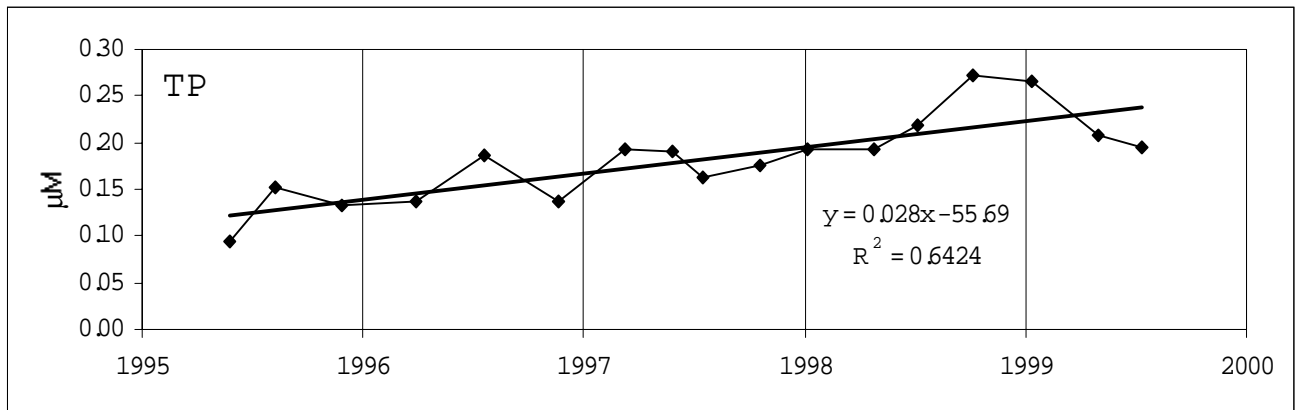
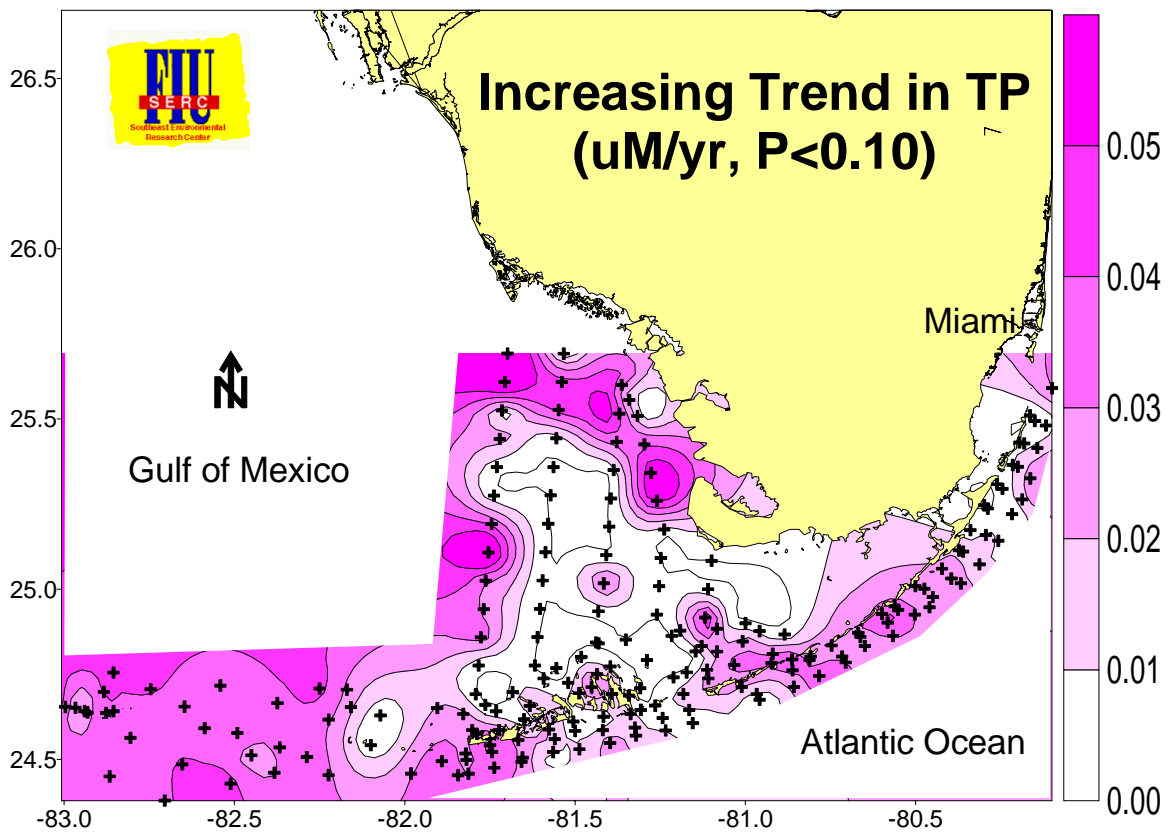## delta Sigma-t
Median for 1995-99

Fig. 2

Fig. 3

Fig. 4

V. List of Appendices

Appendix 1.   Color contour maps of selected water quality variables by sampling event.
These maps encompass all 354 stations of the SERC Water Quality Monitoring
Network which includes the FKNMS, Biscayne Bay, Florida Bay, Whitewater
Bay, Ten Thousand Islands, and Southwest Florida Shelf. The data was collected
over a period of a month so care should be taken in interpreting these maps as
they are not truly synoptic.

Appendix 1 - Contour Maps